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Research review paper

Metabolic reengineering invoked by microbial systems to decontaminate aluminum: Implications for bioremediation technologies

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ABSTRACT

As our reliance on aluminum (Al) increases, so too does its presence in the environment and living systems. Although generally recognized as safe, its interactions with most living systems have been nefarious. This review presents an overview of the noxious effects of Al and how a subset of microbes can rework their metabolic pathways in order to survive an Al-contaminated environment. For instance, in order to expulse metal as an insoluble precipitate, *Pseudomonas fluorescens* shuttles metabolites toward the production of organic acids and lipids that play key roles in chelating, immobilizing and exuding Al. Further, the reconfiguration of metabolic modules enables the microorganism to combat the dearth of iron (Fe) and the excess of reactive oxygen species (ROS) promoted by Al toxicity. While in *Rhizobium* spp., exopolysaccharides have been invoked to sequester this metal, an ATPase is known to safeguard *Anoxybacillus gonensis* against the trivalent metal. Hydroxyl, carboxyl and phosphate moieties have also been exploited by microbes to trap Al. Hence, an understanding of the metabolic networks that are operative in microorganisms residing in polluted environments is critical in devising bioremediation technologies aimed at managing metal wastes. Metabolic engineering is essential in elaborating effective biotechnological processes to decontaminate metal-polluted surroundings.

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Abbreviations: ACN, aconitase; AGODH, acylating glyoxylate dehydrogenase; FUM, fumarase; G6PDH, glucose-6-phosphate dehydrogenase; ICDH, isocitrate dehydrogenase; ICL, isocitrate lyase; KGDH, alpha-ketoglutarate dehydrogenase; MDH, malate dehydrogenase; ME, malic enzyme; NADK, NAD kinase; NOX, NADH oxidase; OCT, oxylate-CoA transferase; PC, pyruvate carboxylase; PE, phosphatidylethanolamine; PEPCK, phosphoenolpyruvate carboxykinase; ROS, reactive oxygen species; SCS, succinyl-CoA synthase; SDH, succinate dehydrogenase.

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1. Introduction

Metal pollution is a threat that all organisms must contend with, particularly in an increasingly industrialized environment. The adverse effects of metal toxicity have been well documented. Heavy metals such as mercury, lead and cadmium as well as toxic metals like Al and chromium in addition to metalloids like arsenic are known to cause major damage to the ecosystem (Bernhoft, 2012; Godt et al.,

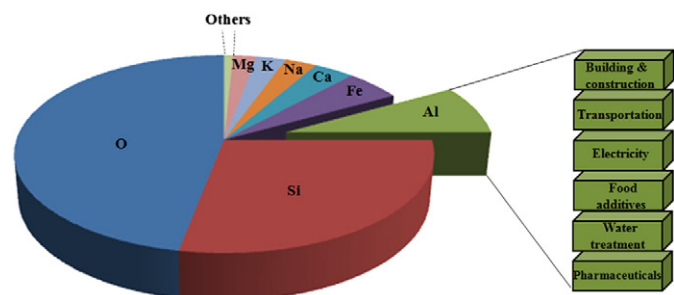


Fig. 1. Elemental composition of the earth's crust and the industrial applications of Al.

2006; Jomova et al., 2011; Patrick, 2006). Given its prevalence in most aspects of our lifestyle, the noxious effects of Al are now beginning to emerge. This trivalent metal is the most abundant metallic element in the earth's crust yet, despite its ubiquity, biological systems have seemingly circumvented the incorporation of this metal in their processes (Verstraeten et al., 2008). The apparent exclusion of Al from all living systems certainly raises questions about its toxicological significance. In spite of an expanding body of literature detailing the harmful consequences of overexposure to Al, there is still substantial debate regarding its safety and usage (Mailloux et al., 2011; Tomljenovic, 2011). Although Al is plentiful, the free metal is almost non-existent, and much of this element exists in complexes which make it inaccessible to most organisms. The increased bioavailability of this metal correlates well with increased anthropogenic activity and utilization (Verstraeten et al., 2008).

The desirability of Al lies in its chemical properties. Being a light, corrosion-resistant material, Al and its alloys have gained considerable traction in numerous fields of human progress (Fig. 1). As the most widely used non-ferrous metal, primary Al production in 2011 was 43.4 million tons, a number that is only predicted to increase in coming years (International Aluminum Institute, 2012). Transportation, particularly in the aerospace industry, benefits greatly from Al's high strength to weight ratio. Furthermore, Al's superconductive properties render it an ideal lightweight alternative to copper wiring. The increase in industrialization has brought about a higher frequency of acid rain, thus lowering the pH of soil and allowing Al to leach into groundwater. This leads to the bioaccumulation of the metal in vegetation and livestock (Piña and Cervantes, 1996). Al's industrial applications notwithstanding, the majority of the exposure to this metal for the average individual comes from diet and drinking water. Certain medications, such as antacid tablets, contain several hundred milligrams of aluminum hydroxide. Additionally, Al is a popular food additive in the form of sodium aluminum phosphates (SALPs). The latter is generally added as emulsifying salts to processed cheese and cheese spreads (López et al., 2002; Yokel et al., 2008). Al is often used as a coagulant in the form of aluminum sulfate (Al₂(SO₄)₃) or polyaluminum chloride (PACl). The application of either compound in drinking water treatment allows for the removal of particulate and dissolved contaminants. However, the end result of this process is usually a higher concentration of Al than was in the water initially (Soni et al., 2001). The average daily human intake of Al from all sources combined ranges between 6 and 14 mg, with variations stemming from age, sex, diet and location (López et al., 2002).

2. Toxicopathology of Al and adaptation to this trivalent metal 107

2.1. Al toxicity 108

Once in the body, Al is distributed and accumulated, particularly in the kidneys, liver and brain. Indeed, this trivalent metal has been linked to several disorders in human beings (Exley and Esiri, 2006; Exley et al., 2006; Lemire and Appanna, 2011; Mailloux et al., 2011; Nayak, 2002). While no single mechanism has been identified which accounts for Al's neurotoxic effects, there is a body of literature detailing its pro-oxidant properties in a number of species (Esparza et al., 2003; Exley, 2004). At concentrations exceeding 0.1 mg/mL in acidic waters (pH 5.0–5.5), Al is acutely toxic to aquatic fauna. Al ions accumulate on and subsequently clog the gill, thus bringing about respiratory dysfunction and accelerating cell necrosis in the gill epithelium (Exley et al., 1991). The induction of oxidative stress and disruption of antioxidant enzymes and neurotransmitter synthesis are apparent in the grass carp *Ctenopharyngodon idella* when exposed to an environment containing 0.1 mg/L Al (Fernández-Dávila et al., 2012). Furthermore, it has been observed that the bioaccumulation of Al in fish can potentially cause damaging effects higher in the food chain (Oberholster et al., 2012).

Research dealing with the exposure of Al in microorganisms has further elaborated on the mechanisms underlying the toxicity of this metal. Concentrations of Al lower than 3 mM have been shown to inhibit the growth of *Escherichia coli*. However, sensitivity to this metal increased with decreasing pH (Guida et al., 1991). Furthermore, the uptake of Al is known to utilize Fe transport systems, thus interfering with the ability of this microbe to capture this essential micronutrient (Davis et al., 1971; El Hage Chahine et al., 2012). This trivalent metal has also been shown to regulate bacterial motility, disrupt plant nodulation and perturb photosynthesis and nitrogen fixation in bacteria (Appanna, 1989; Guzzo et al., 1991; Munns, 1986; Pettersson et al., 1985). Although the molecular details underlying the toxicity of this metal have not been fully delineated, it is clear that the remediation of Al in polluted ecosystems is of paramount importance (Table 1).

2.2. Al-mediated disruption of cell function 141

While Al itself has no role in biological processes, it interferes with metals that do, such as Ca, Mg and Fe (Fig. 2). Al has been shown to increase the intracellular concentrations of Ca, thus perturbing Ca-dependent pathways and contributing to neurotoxicity (Mundy et al., 1997). Mg is known to stabilize phosphate groups in enzymatic reactions and on nucleic acids. Al has been shown to bind to ATP, a key binding partner of Mg, 10⁷ times more tightly than Mg itself (Piña and Cervantes, 1996). Indeed, this complex with ATP has been implicated in the progression of neurodegenerative disease (Exley, 2012). The ability of the Al cation to compete with Mg promotes the loss of membrane fluidity and the disruption of replicative and transcriptional processes in the cell (Johnson and Wood, 1990). As Al and Fe have similar ionic properties, they compete for Fe binding sites in biomolecules. However, while Fe is redox active, Al is not, and the substitution inactivates the target molecule (Middaugh et al., 2005). As we shall see later, this perturbation disables central metabolism and pressures the organism to adapt if it is to survive. Once displaced, labile Fe becomes

Table 1 Examples of locations where the concentration of Al exceeds the maximum permissible limit as defined by regional guidelines.

Region	[Al]	Maximum permissible [all] limit	Source
Gironde/Dordogne, France	Up to 0.514 mg/L	0.1 mg/L	Rondeau, V et al. 2009
Biga Peninsula, Turkey	Up to 15.7 mg/L	0.2 mg/L	Bakar, C et al. 2009
Xi'an city, China	Up to 2.773 mg/L	0.1 mg/L	Wang, D et al. 2012
Chimaliapan wetland, Mexico	Up to 24.82 mg/L	0.02 mg/L	Garcia, G et al. 2012
Maresh and Luda Yana rivers, Bulgaria	Up to 110 mg/L	0.1 mg/L	Rabadjieva, D et al. 2009
Ogun State, Nigeria	Up to 14 mg/L	0.2 mg/L	Odukoya, O et al. 2010

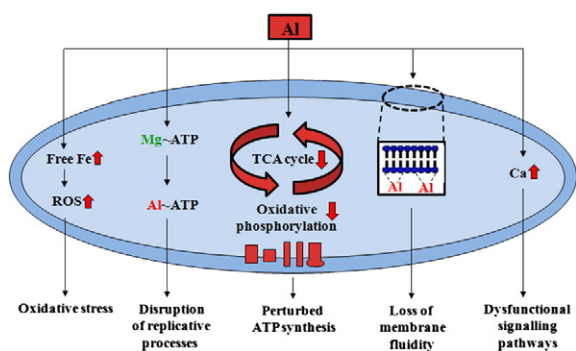


Fig. 2. An overview of the cellular processes affected by an Al-rich microenvironment.

a potent generator of hydroxyl radicals via the Fenton reaction, thus creating oxidative stress and exacerbating the burden brought about by Al (Mailloux et al., 2011). Additionally, this metal is known to interfere with protein kinase C, cAMP homeostasis and glutamate-nitric oxide synthase-cGMP signaling networks (Hartle et al., 1996; Lajeunesse et al., 1998; Quarles et al., 1994).

2.3. Adaptation to Al

To date, most studies on Al tolerance have been performed on plants, where the interaction of this metal with the root apex in acidic soils has been linked to the reduction of crop yields (Krill et al., 2010). The activation of anion transporters in the plasma membrane permit the plant to exude organic acids and phenolic compounds which chelate Al and negate its toxicity (Krill et al., 2010). Several species of fungi are known to adopt a similar strategy to dispose of this trivalent metal, utilizing negatively charged metabolites such as citrate and malate to chelate and reduce free Al (Klugh and Cumming, 2007). Indeed, such a strategy limits the bioavailability of the metal in the vicinity of plant roots, thus conferring resistance to Al's noxious effects (Kelly et al., 2005).

Mechanisms of tolerance tailored specifically to this metal are scarcely found in microbial systems. In general, chelation by organic acids, followed by exudation or compartmentalization is commonly utilized to control Al's dangerous effects. However, it has recently been discovered that the gene product of G2ALT from the thermophilic bacterium *Anoxybacillus gonensis* is a PP-loop type ATPase which plays a role in Al expulsion in the microbe. Indeed, G2ALT has an 87% similarity with alu1-p, a gene associated with the Al tolerance phenotype from *Actinomyces viscosus* (Beris et al., 2011). Reports of a community phylogenetically related to the dissimilatory-sulfate reducing bacteria *Desulfovibrio desulfuricans* describe their ability to remove 85% of Al in a 0.9 mM solution in the presence of the sulfate anion. X-ray diffraction (XRD) and transmission electron microscopy (TEM-EDS) were utilized to uncover the precipitation of this metal in an amorphous Al-hydroxide aggregate after 27 days of incubation (Martins et al., 2012).

The combination of clay and strain MAM-4 of *Providencia rettgeri* has been applied to the removal of Al, Cu and Co from wastewater. Indeed, the negatively charged clay in conjunction with the ability of *P. rettgeri* to withstand high concentrations (> 18 mM) of Al permitted the removal of 87% of this trivalent metal. Fourier transform infrared spectroscopy (FTIR) delineated that hydroxyl, carboxyl and phosphate groups were applied by the microbe for the biosorption of these heavy metals (Abo-Amer et al., 2012). In another study, the generation of exopolysaccharides by nitrogen fixing *Rhizobium* strains correlates with resistance to concentrations of Al up to 1 mM in acidic conditions. Additional defense mechanisms in these microbes include the reduction of negative charges in the cell membrane and the formation of insoluble aluminum phosphate complexes (Avelar Ferreira et al., 2012).

Although a plethora of Al-resistant microorganisms have been uncovered, a detailed analysis of this metal's toxic effects on intracellular pathways has eluded researchers. Here, we elaborate on how the soil microbe *Pseudomonas fluorescens* utilizes its metabolic versatility to survive in Al-rich environments. The reengineering of metabolic networks aimed at eliminating Al and fending ROS is described. Furthermore, the significance of these Al-induced metabolic modules in bioremediation technologies is elaborated. It is critical that the molecular details conferring metal resistance be unraveled if microbial-mediated biotechnology aimed at waste metal decontamination is to be optimized.

3. *P. fluorescens*, an Al-resistant microbe

P. fluorescens derives its name from its ability to synthesize the fluorescent siderophore pyoverdinin, and is predominantly a soil bacterium known for its ability to adapt to a great variety of nutrient sources (Paulsen et al., 2005). Metabolic flexibility has rendered this microbe suitable for a variety of biotechnological processes. While this bacterium can survive in temperatures as low as 0 °C, its optimal growth temperature is 25 °C. Its inability to grow at body temperature means *P. fluorescens* is considered non-pathogenic and its presence in the environment not deemed harmful to mammals (Srivastava, 2009). This microbe frequently colonizes the rhizosphere of crop plants where it displays antagonistic activity towards various plant pathogens. Fungi, such as *Alternaria cajani*, *Curvularia lunata* and *Fusarium sp.* which may cause disease and death in the target plants, are eradicated in the presence of *P. fluorescens*. To accomplish this, the organism produces a number of antifungal metabolites, such as 2,4-diacetylphloroglucinol, phenazine-1-carboxylic acid, 2, 3-de-epoxy-2, 3-didehydra-rhizoxin and pyrrolnitrin. The latter has been used as the lead structure for the development of a novel agricultural fungicide. Additionally, *P. fluorescens* produces secondary metabolites such as siderophores, HCN and proteases that directly kill or scavenge nutrients from competing organisms. Indeed, the application of this microbe in biological control offers a safer alternative to chemical control, which engenders environmental concerns (Srivastava, 2009).

We have previously reported on the ability of *P. fluorescens* to withstand concentrations of Al up to 50 mM (Appanna and St. Pierre, 1994). The metal is chelated to citric acid, which translocates across the bacterial membrane as a complex which is then degraded intracellularly. After 80 h of growth, 65.1% of the soluble Al is sequestered when *P. fluorescens* is grown in media containing an Al concentration of 15 mM (Appanna and St. Pierre, 1996). Analysis of the soluble cell free extract following 75 h of growth showed that less than 1% of the Al in the media was found in this compartment (Appanna and St. Pierre, 1996). The ability of this bacterium to cope with such a high concentration of Al involves a detoxification strategy, which consists primarily of excluding it from the cell as an insoluble lipid containing precipitate tailored to alleviate the toxicity of this metal (Appanna and St. Pierre, 1996) (Fig. 3A,B). The nature of this gelatinous residue harboring Al was shown to be phosphatidylethanolamine (PE), a phospholipid commonly found in biological membranes (Appanna and St. Pierre, 1996).

Scanning electron microscopy and transmission electron microscopy (SEM and TEM) were utilized to ascertain the nature of this precipitate, and the percentage of Al found sequestered ranged from 38 to 80% in cultures with 1 to 30 mM Al (Appanna and St. Pierre, 1996). In addition, the amount of PE found in the spent fluid increased with increasing concentrations of Al, thus confirming that this phospholipid is intricately linked to the detoxification of this trivalent metal. Indeed, the addition of cerulenin, an inhibitor of lipid synthesis, greatly limits the amount of Al removed (Hamel and Appanna, 2003). Clearly, the exudation of Al is the primary stratagem by which the microbe copes with its toxicity. This is true of other metals, such as gallium (Ga) and indium (In), where disparate stratagems are deployed.

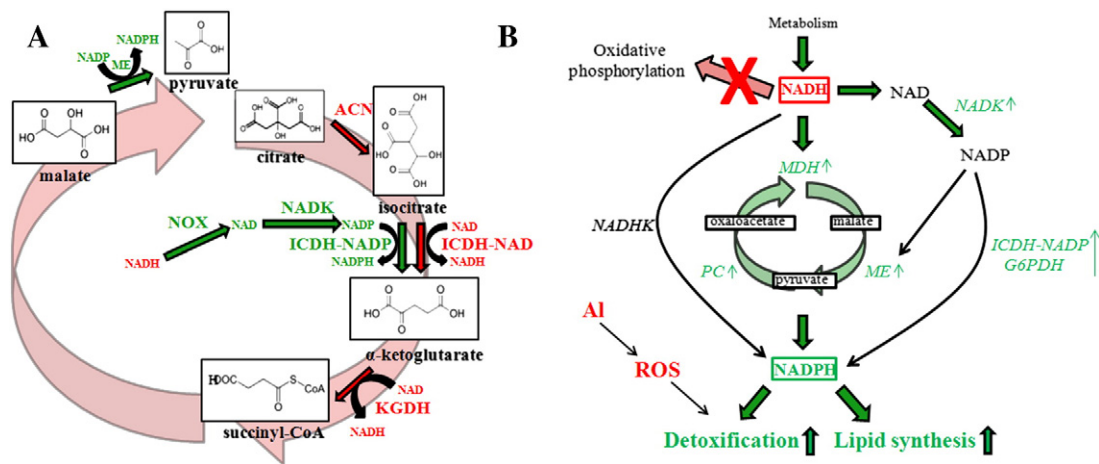


Fig. 4. Metabolic networks to decontaminate Al in *P. fluorescens*. A: Al triggers a metabolic shift that favors the synthesis of NADPH and curbs the intracellular concentration of the pro-oxidant NADH in *P. fluorescens*. B: Metabolic reengineering aimed at the conversion of NADH to NADPH allows for the detoxification of Al-derived ROS and the synthesis of lipid moieties for exudation of this trivalent metal. ICDH: isocitrate dehydrogenase; KGDH: α -ketoglutarate dehydrogenase; ME: malic enzyme; NOX: NADH oxidase; NADK: NAD kinase; MDH: malate dehydrogenase; PC: pyruvate carboxylase; ME: malic enzyme; NADHK: NADH kinase; G6PDH: glucose-6-phosphate dehydrogenase. Adapted from Mailloux et al., 2008.

further limit their activity and maintain a reductive environment, *P. fluorescens* decreases the intracellular concentration of NADH. Two stratagems are invoked to accomplish this. Firstly, NADH producing enzymes, such as ICDH-NAD and α -ketoglutarate dehydrogenase (KGDH) are down-regulated (Mailloux et al., 2007). These enzymes are not required in the modified TCA cycle utilized by this microbe as they generate NADH (Singh et al., 2009). Additionally, the increased expression of ICDH-NADP accompanied by the lowered activity of KGDH builds a pool of α -ketoglutarate, a keto-acid which has been shown to scavenge ROS in situ (Varma and Hegde, 2004). Secondly, the direct consumption of NADH is encouraged by the microbe via the expression of H₂O-generating NADH oxidase (NOX) (Chenier et al., 2008). The product of this reaction, NAD, can be utilized to propel redox reactions or for NADP generation by NADK. The enhanced production of NADPH does not only help fend Al-induced oxidative stress, but it also provides the crucial reducing factor required for lipogenesis, an event that results in the synthesis of PE. The latter contributes to the extrusion of the contaminant from the microbial system. Furthermore, *P. fluorescens* elaborates a metabolic circuit that results in the conversion of NADH into NADPH. In this instance, oxaloacetate is converted to pyruvate with the concomitant production of NADPH. Malate dehydrogenase (MDH), ME and pyruvate carboxylase mediate this process (Fig. 4B).

3.3. Organic acids as potent aluminophores

While PE synthesis offers one route of Al expulsion, it is not the only moiety on which *P. fluorescens* relies for the elimination of this metal. Organic acids, owing to their negatively charged carboxylic acid groups, are well-suited for the chelation and removal of noxious heavy metals (Ma et al., 2001). In order to ameliorate the biogenesis of these aluminophores, *P. fluorescens* undergoes a complete overhaul of its metabolic networks. The first of these changes is the increased expression of OCT. This enzyme permits the formation of succinyl-CoA from oxalyl-CoA (Singh et al., 2009). While the former can be utilized to produce ATP from SLP as described earlier, the latter becomes oxalate, a dicarboxylic acid and potent chelator of Al (Hamel and Appanna, 2001). Thus, while generating energy, the microbe finds an effective route to the production of oxalate. This reengineered metabolic network allows the bacterium to produce both ATP and a chelator of Al.

A second metabolic reconfiguration occurs in order to generate another dicarboxylic acid, oxaloacetate, which is used as a precursor

to a polycarboxylic aluminophore (Appanna et al., 2003; Lemire et al., 2008). The homeostasis of this metabolite is regulated by numerous enzymes. Malate dehydrogenase (MDH), a critical TCA cycle enzyme, appears to be up-regulated by the microbe in order to produce oxaloacetate at a greater rate when Al is present (Lemire et al., 2008). Pyruvate carboxylase (PC), which carboxylates pyruvate to form oxaloacetate, is also overexpressed by *P. fluorescens* treated with Al (Lemire et al., 2008). On the other hand, phosphoenolpyruvate carboxykinase (PEPCK), an enzyme that converts oxaloacetate to phosphoenolpyruvate shows lower expression (Lemire et al., 2008). Furthermore, the activity of nucleoside diphosphate kinase (NDPK), a critical enzyme for the generation of the PEPCK cofactor GTP, also shows lowered activity in the Al-treated bacteria (Lemire et al., 2008). PC and PEPCK are normally coupled in the gluconeogenic pathway, allowing the organism to produce intermediates for biosynthetic reactions (Owen et al., 2002). However, it appears as though Al toxicity disrupts this process.

These metabolic shifts aimed at the production and pooling of two dicarboxylic acids, oxaloacetate and oxalate, permit the microbe to expulse Al as a PE containing precipitate and continue to proliferate (Fig. 5). Hence, the metabolic reengineering invoked to neutralize the toxic effects of Al enables the production of NADPH, ATP, oxaloacetate, oxalate, and PE. While the latter three molecules are directly involved in the immobilization of the trivalent metal, NADPH provides the redox environment to sustain life in an oxidatively charged milieu and the reducing equivalent for lipid biosynthesis. Despite the diminished effectiveness of oxidative phosphorylation at fulfilling the need for high-energy phosphate under Al assault, the microbe devises a unique metabolic route to ATP and oxalate that is devoid of the demand of oxygen. These reengineered metabolic networks are keys to the survival of *P. fluorescens* in an Al-contaminated environment and in the elimination of the toxic metal. Thus, understanding these metabolic changes in Al-tolerant bacteria is essential in bioremediation technologies.

4. Proposed bioremediation device fabrication and applications

As high Al levels (≥ 0.1 mg/L) in drinking water have been associated with an elevated risk of dementia and Alzheimer's disease, the removal of this heavy metal in contaminated regions may alleviate this burden (Rondeau et al., 2000). Biotreatment of wastewater has been shown to efficiently remove a number of organic contaminants such as hydrocarbons (oil), polychlorinated biphenyls (PCBs) and pharmaceutical substances (Bao et al., 2012; Federici et al., 2012;

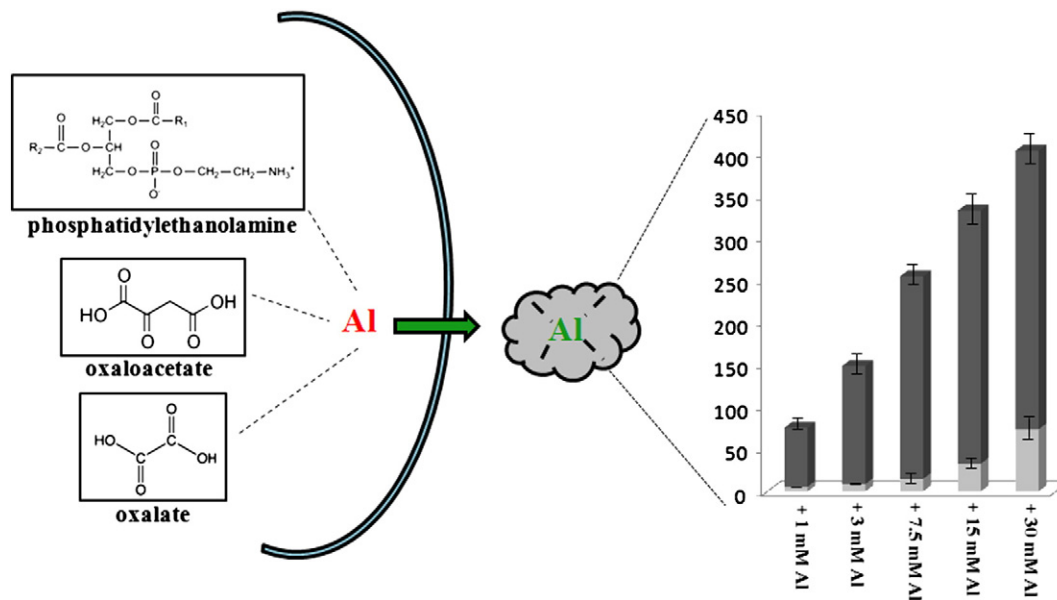


Fig. 5. Under Al stress, *P. fluorescens* favors the biosynthesis of phosphatidylethanolamine and dicarboxylic acids which bind to and expulse Al as an insoluble pellet. Pellet weight and Al content vary with the concentration of this metal in the medium. □; pellet weight (mg/mL culture). ■; aluminum (nmol/mg [wet wt] of pellet).

422 Mansour et al., 2012). However, inorganic waste such as metals can-
 423 not be degraded into benign compounds like H₂O and CO₂. The ability
 424 of microorganisms to alter the speciation of metals, leading to their
 425 bioprecipitation, volatilization, or biotransformation into innocuous
 426 forms has become a key tool in environmental remediation (Singh
 427 and Cameotra, 2004). Bioprecipitation concentrates the metals and
 428 thus reduces their bioavailability. Further, these biotransformed con-
 429 centrates can be stored in backfills or by vitrification. Several physical
 430 and chemical methods are currently undertaken to remediate Al and
 431 other metals in polluted ecosystems. These include chemical precipi-
 432 tation or reduction, ion-exchange, activated charcoal, adsorption and
 433 reverse osmosis among others (Srivastava and Majumder, 2008).
 434 However, the cost of reagents and materials in addition to the possi-
 435 bility of secondary environmental pollution become limiting factors
 436 for these treatments (Malik, 2004).

437 Given that *P. fluorescens* can tolerate such high concentrations of
 438 Al while simultaneously exuding it as an insoluble precipitate, this
 439 microbe is well-suited for the removal of this metal from polluted
 440 ecosystems. Therefore, we propose that *P. fluorescens* be applied in
 441 the fabrication of a biofilter (Fig. 6). The latter may incorporate the
 442 fixing of the microorganism to a porous medium to decontaminate
 443 organic and inorganic contaminants from wastewater (Srivastava
 444 and Majumder, 2008). By immobilizing the growing biofilm in a ro-
 445 tating or fixed reactor, the waste stream containing dissolved Al

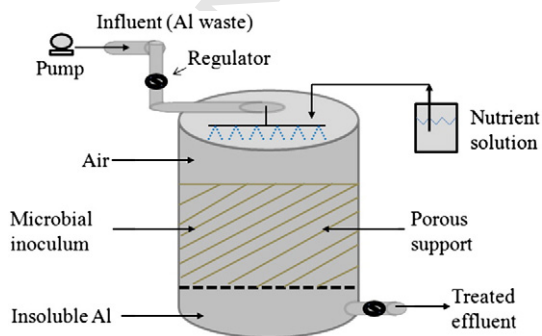


Fig. 6. Proposed bioreactor design. A trickling biofilter which feeds in the wastewater and a nutrient source for the microbe permits the bioprecipitation of Al, which is filtered out of the system. Valves controlling intake and outtake determine the flow rate, allowing for maximal removal of the trivalent metal.

446 can be treated and the metal subsequently filtered out. The vessel
 447 would function at ambient temperature and a pH ranging from mildly
 448 acidic to neutral. Rotation speed and flow rate are parameters which
 449 can be optimized for Al removal. A nutrient broth, supplied daily via a
 450 peristaltic pump may be utilized to support the viability of the microbe.
 451 Due to the metabolic versatility of *P. fluorescens*, multiple carbon
 452 sources can be utilized for sustenance. These Al-resistant bacteria can
 453 also be maintained in a defined mineral medium consisting of 6.0 g
 454 Na₂HPO₄, 3.0 g KH₂PO₄, 0.8 g NH₄Cl, 0.2 g MgSO₄ and 4.0 g citrate
 455 per L of media. Trace metals consisting of 2 μM Fe, 1 μM Mn, 0.5 μM
 456 Zn, 1 μM Ca, 0.25 μM Co, 0.1 μM Cu and 0.1 μM Mo may help in the pro-
 457 cess. Precipitated Al is eliminated at the bottom of the receptacle via a
 458 filter while the decontaminated water is collected for use and/or re-
 459 leased into the environment indeed after verifying the Al and microbial
 460 content. Although *P. fluorescens* is being exploited as a model microbe
 461 in the foregoing progress, other Al-resistant microbes can also be
 462 applied in the fabrication of biofilters, either alone or in the form of a
 463 mixed biofilm. Indeed, heterogeneous bacterial communities often re-
 464 move the undesired contaminant more effectively than a single species
 465 (Srivastava and Majumder, 2008).

466 Alternatively, these microbes can be utilized in bioaugmentation,
 467 as seeding organisms which precipitate aluminum from the soil. In
 468 effect, *Pseudomonas* inoculants have been applied with great success
 469 to protect plants from noxious chemicals and organisms (Ajithkumar
 470 et al., 1998; Li and Alexander, 1988). Inoculation of Al-contaminated
 471 soil with *P. fluorescens* may precipitate this metal, thus rendering it
 472 unavailable for uptake in plants and augmenting crop yield.

473 Although the pathogenesis underlying most neurodegenerative
 474 disorders is multifactorial, one cannot discount the increasing body
 475 of evidence that high concentrations of Al are associated with a variety
 476 of diseased states. A number of regions around the globe are laden
 477 with quantities of this metal which are beyond the maximum permis-
 478 sible limit (Table 1). The removal of Al from contaminated waters
 479 may have a positive effect on human health, crop viability and the bal-
 480 ance of the ecosystem. While purification strategies do exist, many of
 481 these procedures are inaccessible due to economic reasons, particu-
 482 larly in less developed nations (Bakar et al., 2010). Bioremediation
 483 of wastewater is both cost-effective and benign, particularly when
 484 compared to alternatives such as chemical treatment or incineration
 485 which entail risks to the environment (Bitton and Koopman, 1992).
 486 *P. fluorescens*' resistance to Al renders it an ideal microorganism for

487 bioremediation technologies. While it is feasible that genetic alter-
488 ations would enhance the ability of this microorganism to dispose
489 of Al, one of the primary advantages of this system is its innateness;
490 i.e. no genetic manipulation is involved. Indeed, the introduction of
491 genetically modified organisms into the environment has long since
492 had a negative stigma attached to it (Marris, 2001).

493 5. Conclusion

494 Here we demonstrated how microorganisms can reengineer cellular
495 processes to bioprecipitate Al, thus rendering it unavailable for biologi-
496 cal uptake in other organisms. The metabolic modules that render this
497 decontamination strategy feasible allow for the biotransformation and
498 expulsion of Al without significantly impeding the survival of the organ-
499 ism. The natural resistance to high concentrations of Al demonstrated
500 by certain microbes is a feature which can be harnessed to design biore-
501 actors and the molecular details outlined may indeed help improve the
502 efficiency of such processes. Metabolic engineering is pivotal in elabo-
503 rating efficient biotechnological processes aimed at decontaminating
504 metal pollutants.

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512 References

- 513 Abo-Amer AE, Ramadan AB, Abo-State M, Abu-Ghabia MA, Ahmed HE. Biosorption of
514 aluminum, cobalt and copper ions by *Providencia rettgeri* isolated from wastewater.
515 J Basic Microbiol 2012;10:00635.
516 Ajithkumar PV, Gangadhara KP, Manilal P, Kunhi AAM. Soil inoculation with *Pseudomonas*
517 *aeruginosa* 3mT eliminates the inhibitory effect of 3-chloro- and 4-chlorobenzoate on
518 tomato seed germination. Soil Biol Biochem 1998;30:1053–9.
519 Anderson S, Appanna VD. Indium detoxification in *Pseudomonas fluorescens*. Environ
520 Pollut 1993;82:33–7.
521 Appanna VD. Exopolysaccharide syntheses in *Rhizobium meliloti* in the presence of
522 manganese and aluminum. Microbios Lett 1989;40:31–6.
523 Appanna VD, St. Pierre M. Influence of phosphate on aluminum tolerance in *Pseudomonas*
524 *fluorescens*. FEMS Microbiol Lett 1994;124:327–32.
525 Appanna VD, St. Pierre M. Aluminum elicits exocellular phosphatidylethanolamine
526 production in *Pseudomonas fluorescens*. Appl Environ Microbiol 1996;62:2778–82.
527 Appanna VD, Gazsó LG, St. Pierre M. Multiple-metal tolerance in *Pseudomonas*
528 *fluorescens* and its biotechnological significance. J Biotechnol 1996;52:75–80.
529 Appanna VD, Hamel R, Mackenzie C, Kumar P, Kalyuzhnyi SV. Adaptation of *Pseudomonas*
530 *fluorescens*; to Al-citrate: involvement of tricarboxylic acid and glyoxylate cycle
531 enzymes and the influence of phosphate. Curr Microbiol 2003;47:521–7.
532 Avelar Ferreira PA, Bomfeti CA, Lima Soares B, de Souza Moreira FM. Efficient
533 nitrogen-fixing *Rhizobium* strains isolated from amazonian soils are highly tolerant
534 to acidity and aluminium. World J Microbiol Biotechnol 2012;28:1947–59.
535 Bakar C, Karaman HI, Baba A, Sengunalp F. Effect of high aluminum concentration in
536 water resources on human health, case study: Biga Peninsula, northwest part of
537 Turkey. Arch Environ Contam Toxicol 2010;58:935–44.
538 Bao MT, Wang LN, Sun PY, Cao LX, Zou J, Li YM. Biodegradation of crude oil using an
539 efficient microbial consortium in a simulated marine environment. Mar Pollut
540 Bull 2012;64:1177–85.
541 Bériault R, Hamel R, Chenier D, Mailloux R, Joly H, Appanna V. The overexpression of
542 NADPH-producing enzymes counters the oxidative stress evoked by gallium, an
543 iron mimetic. Biometals 2007;20:165–76.
544 Beris FS, De Smet L, Karaoglu H, Canakci S, Van Beeumen J, Belduz AO. The ATPase
545 activity of the G2alt gene encoding an aluminium tolerance protein from
546 *Anoxybacillus gonensis* G2. J Microbiol 2011;49:641–50.
547 Bernhoft RA. Mercury toxicity and treatment: a review of the literature. J Environ Public
548 Health 2012;460508.
549 Bitton G, Koopman B. Bacterial and enzymatic bioassays for toxicity testing in the
550 environment. Rev Environ Contam Toxicol 1992;125:1–22.
551 Bochud-Allemann N, Schneider A. Mitochondrial substrate level phosphorylation is
552 essential for growth of procyclic *Trypanosoma brucei*. J Biol Chem 2002;277:
553 32849–54.
554 Chenier D, Bériault R, Mailloux R, Baquie M, Abramia G, Lemire J, et al. Involvement of
555 fumarase C and NADH oxidase in metabolic adaptation of *Pseudomonas fluorescens*

- cells evoked by aluminum and gallium toxicity. Appl Environ Microbiol 2008;74: 556
3977–84. 557
Davis WB, McCauley MJ, Byers BR. Iron requirements and aluminum sensitivity of an 558
hydroxamic acid-requiring strain of *Bacillus megaterium*. J Bacteriol 1971;105: 559
589–94. 560
El Hage Chahine JM, Hemadi M, Ha-Duong NT. Uptake and release of metal ions by trans- 561
ferrin and interaction with receptor 1. Biochim Biophys Acta 2012;1820:334–47. 562
Esparza JL, Gómez M, Romeu M, Mulero M, Sánchez DJ, Mallol J, et al. 563
Aluminum-induced pro-oxidant effects in rats: protective role of exogenous mel- 564
atonin. J Pineal Res 2003;35:32–9. 565
Exley C. The pro-oxidant activity of aluminum. Free Radic Biol Med 2004;36:380–7. 566
Exley C. The coordination chemistry of aluminium in neurodegenerative disease. Coord 567
Chem Rev 2012;256:2142–6. 568
Exley C, Esiri MM. Severe cerebral congophilic angiopathy coincident with increased 569
brain aluminium in a resident of Camelford, Cornwall, UK. J Neurol Neurosurg 570
Psychiatry 2006;77:877–9. 571
Exley C, Chappell JS, Birchall JD. A mechanism for acute aluminium toxicity in fish. 572
J Theor Biol 1991;151:417–28. 573
Exley C, Begum A, Woolley MP, Bloor RN. Aluminum in tobacco and cannabis and 574
smoking-related disease. Am J Med 2006;119(276):e9–11. 575
Federici E, Giubilei M, Santi G, Zanolari G, Negroni A, Fava F, et al. Bioaugmentation of 576
a historically contaminated soil by polychlorinated biphenyls with *Lentinus tigrinus*. 577
Microb Cell Fact 2012;11:35. 578
Fernández-Dávila ML, Razo-Estrada AC, García-Medina S, Gómez-Oliván LM, 579
Piñón-López MJ, Ibarra RG, et al. Aluminum-induced oxidative stress and neuro- 580
toxicity in grass carp (Cyprinidae – *Ctenopharingodon idella*). Ecotoxicol Environ 581
Saf 2012;76:87–92. 582
Freinbichler W, Colivicchi M, Stefanini C, Bianchi L, Ballini C, Misini B, et al. Highly 583
reactive oxygen species: detection, formation, and possible functions. Cell Mol 584
Life Sci 2011;68:2067–79. 585
Godt J, Scheidig F, Grosse-Siestrup C, Esche V, Brandenburg P, Reich A, et al. The toxicity 586
of cadmium and resulting hazards for human health. J Occup Med Toxicol 2006;1: 587
22. 588
Grose JH, Joss L, Velick SF, Roth JR. Evidence that feedback inhibition of NAD kinase 589
controls responses to oxidative stress. Proc Natl Acad Sci 2006;103:7601–6. 590
Guida L, Saidi Z, Hughes MN, Poole RK. Aluminium toxicity and binding to *Escherichia* 591
coli. Arch Microbiol 1991;156:507–12. 592
Guzzo A, Diorio C, DuBow MS. Transcription of the *Escherichia coli* *fliC* gene is regulated 593
by metal ions. Appl Environ Microbiol 1991;57:2255–9. 594
Hamel RD, Appanna VD. Modulation of TCA cycle enzymes and aluminum stress in 595
Pseudomonas fluorescens. J Inorg Biochem 2001;87:1–8. 596
Hamel R, Appanna VD. Aluminum detoxification in *Pseudomonas fluorescens* is medi- 597
ated by oxalate and phosphatidylethanolamine. Biochim Biophys Acta 2003;1619: 598
70–6. 599
Hartle JE, Prcic V, Siddhanti SR, Spurney RF, Quarles LD. Differential regulation of 600
receptor-stimulated cyclic adenosine monophosphate production by polyvalent 601
cations in MC3T3-E1 osteoblasts. J Bone Miner Res 1996;11:789–99. 602
Johnson AC, Wood M. DNA, a possible site of action of aluminum in *Rhizobium* spp. 603
Appl Environ Microbiol 1990;56:3629–33. 604
Jomova K, Jenisova Z, Feszterova M, Baros S, Liska J, Hudecova D, et al. Arsenic: toxicity, 605
oxidative stress and human disease. J Appl Toxicol 2011;31:95–107. 606
Kelly CN, Morton JB, Cumming JR. Variation in aluminum resistance among arbuscular 607
mycorrhizal fungi. Mycorrhiza 2005;15:193–201. 608
Klugh KR, Cumming JR. Variations in organic acid exudation and aluminum resistance 609
among arbuscular mycorrhizal species colonizing *Liriodendron tulipifera*. Tree 610
Physiol 2007;27:1103–12. 611
Krill AM, Kirst M, Kochian LV, Buckler ES, Hoekenga OA. Association and linkage anal- 612
ysis of aluminum tolerance genes in maize. PLoS One 2010;5:e9958. 613
Kussmaul L, Hirst J. The mechanism of superoxide production by NADH:ubiquinone 614
oxidoreductase (complex I) from bovine heart mitochondria. Proc Natl Acad Sci 615
2006;103:7607–12. 616
Lajeunesse D, Moreau R, Hobbs W, Qui W, Lafond J, Guggino. Influence of aluminum on 617
the regulation of PTH- and 1,25(OH)2D3-dependent pathways in the rat osteosar- 618
coma cell line ROS 17/2.8. J Bone Miner Res 1998;13:962–9. 619
Lemire J, Appanna VD. Aluminum toxicity and astrocyte dysfunction: a metabolic link 620
to neurological disorders. J Inorg Biochem 2011;105:1513–7. 621
Lemire J, Kumar P, Mailloux R, Cossar K, Appanna VD. Metabolic adaptation and oxalo- 622
acetate homeostasis in *P. fluorescens* exposed to aluminum toxicity. J Basic 623
Microbiol 2008;48:252–9. 624
Lemire J, Mailloux R, Auger C, Whalen D, Appanna VD. *Pseudomonas fluorescens* orches- 625
trates a fine metabolic-balancing act to counter aluminium toxicity. Environ 626
Microbiol 2010;12:1384–90. 627
Li DM, Alexander M. Co-inoculation with antibiotic-producing bacteria to increase 628
colonization and nodulation by rhizobia. Plant and Soil 1988;108:211–9. 629
López FF, Cabrera C, Lorenzo ML, López MC. Aluminum levels in convenience and fast 630
foods: in vitro study of the absorbable fraction. Sci Total Environ 2002;300:69–79. 631
Ma JF, Ryan PR, Delhaize E. Aluminium tolerance in plants and the complexing role of 632
organic acids. Trends Plant Sci 2001;6:273–8. 633
Mailloux RJ, Bériault R, Lemire J, Singh R, Chénier DR, Hamel RD, et al. The tricarboxylic 634
acid cycle, an ancient metabolic network with a novel twist. PLoS One 2007;2: 635
e690. 636
Mailloux RJ, Lemire J, Appanna VD. Hepatic response to aluminum toxicity: 637
dyslipidemia and liver diseases. Exp Cell Res 2011;317:2231–8. 638
Malik A. Metal bioremediation through growing cells. Environ Int 2004;30:261–78. 639
Mansour BH, Mosrati R, Barillier D, Ghedira K, Chekir-Ghedira L. Bioremediation of 640
industrial pharmaceutical drugs. Drug Chem Toxicol 2012;35:235–40. 641

- 642 Marris C. Public views on GMOs: deconstructing the myths. Stakeholders in the GMO
643 debate often describe public opinion as irrational. But do they really understand
644 the public? *EMBO Rep* 2001;2:545–8.
- 645 Martins M, Taborda R, Silva G, Assunção A, Matos AP, Costa MC. Aluminum and removal by
646 a highly Al-resistant dissimilatory sulphate-reducing bacteria community. *Biodegrada-*
647 *tion* 2012;23:693–703.
- 648 Middaugh J, Hamel R, Jean-Baptiste G, Beriault R, Chenier D, Appanna VD. Aluminum
649 triggers decreased aconitase activity via Fe-S cluster disruption and the overexpression
650 of isocitrate dehydrogenase and isocitrate lyase: a metabolic network mediating
651 cellular survival. *J Biol Chem* 2005;280:3159–65.
- 652 Mundy W, Freudenrich T, Kodavanti P. Aluminum potentiates glutamate-induced cal-
653 cium accumulation and iron-induced oxygen free radical formation in primary
654 neuronal cultures. *Mol Chem Neurobiol* 1997;32:41–57.
- 655 Munns DN. Acid soil tolerance in legumes and *rhizobia*. *Adv Plant Nutr* 1986;32:63–91.
- 656 Nayak P. Aluminum: impacts and disease. *Environ Res* 2002;89:101–15.
- 657 Oberholster PJ, Myburgh JG, Ashton PJ, Coetzee JJ, Botha AM. Bioaccumulation of alu-
658 minium and iron in the food chain of Lake Loskop, South Africa. *Ecotoxicol Environ*
659 *Saf* 2012;75:134–41.
- 660 Owen OE, Kalhan SC, Hanson RW. The key role of anaplerosis and cataplerosis for citric
661 acid cycle function. *J Biol Chem* 2002;277:30409–12.
- 662 Patrick L. Lead toxicity part II: the role of free radical damage and the use of antioxi-
663 dants in the pathology and treatment of lead toxicity. *Altern Med Rev* 2006;11:
664 114–27.
- 665 Paulsen IT, Press CM, Ravel J, Kobayashi DY, Myers GSA, Mavrodi DV, et al. Complete
666 genome sequence of the plant commensal *Pseudomonas fluorescens* Pf-5. *Nat*
667 *Biotechnol* 2005;23:873–8.
- 668 Pettersson A, Hällbom L, Bergman B. Physiological and structural responses of the cya-
669 nobacterium *Anabaena cylindrica* to aluminium. *Physiol Plant* 1985;63:153–8.
- 670 Piña RG, Cervantes C. Microbial interactions with aluminium. *Biometals* 1996;9:311–6.
- 671 Quarles LD, Hartle JE, Middleton JP, Zhang J, Arthur JM, Raymond JR.
672 Aluminum-induced DNA synthesis in osteoblasts: mediation by a G-protein
673 coupled cation sensing mechanism. *J Cell Biochem* 1994;56:106–17.
- 674 Rondeau V, Commenges D, Jacqmin-Gadda H, Dartigues JF. Relation between aluminum
675 concentrations in drinking water and Alzheimer's disease: an 8-year follow-up
676 study. *Am J Epidemiol* 2000;152:59–66.
- Singh P, Cameotra SS. Enhancement of metal bioremediation by use of microbial
677 surfactants. *Biochem Biophys Res Commun* 2004;319:291–7. 678
- Singh R, Lemire J, Mailloux RJ, Appanna VD. A novel strategy involved in anti-oxidative
679 defense: the conversion of NADH into NADPH by a metabolic network. *PLoS One*
680 2008;3:e2682. 681
- Singh R, Lemire J, Mailloux RJ, Chénier D, Hamel R, Appanna VD. An ATP and oxalate
682 generating variant tricarboxylic acid cycle counters aluminum toxicity in *Pseudomonas*
683 *fluorescens*. *PLoS One* 2009;4:e7344. 684
- Soni MG, White SM, Flamm WG, Burdock GA. Safety evaluation of dietary aluminum. *685*
Regul Toxicol Pharmacol 2001;33:66–79. 686
- Srivastava R. Antifungal activity of *Pseudomonas fluorescens* against different plant path-
687 ogenic fungi. *Internet J Microbiol* 2009;7:2. 688
- Srivastava NK, Majumder CB. Novel biofiltration methods for the treatment of heavy
689 metals from industrial wastewater. *J Hazard Mater* 2008;151:1–8. 690
- Tomljenovic L. Aluminum and Alzheimer's disease: after a century of controversy, is
691 there a plausible link? *J Alzheimers Dis* 2011;23:567–98. 692
- Varma SD, Hegde KR. Effect of α -ketoglutarate against selenite cataract formation. *Exp*
693 *Eye Res* 2004;79:913–8. 694
- Verstraeten S, Aimo L, Oteiza P. Aluminium and lead: molecular mechanisms of brain
695 toxicity. *Arch Toxicol* 2008;82:789–802. 696
- Yamamoto Y, Kobayashi Y, Devi SR, Rikiishi S, Matsumoto H. Aluminum toxicity is
697 associated with mitochondrial dysfunction and the production of reactive oxygen
698 species in plant cells. *Plant Physiol* 2002;128:163–72. 699
- Yokel RA, Hicks CL, Florence RL. Aluminum bioavailability from basic sodium aluminum
700 phosphate, an approved food additive emulsifying agent, incorporated in cheese. *701*
Food Chem Toxicol 2008;46:2261–6. 702

Web references

- International Aluminum Institute. <http://www.world-aluminium.org/Statistics> 2012. 703
[Accessed May 14th.2012]. 704
705
706
707